

Long-term Export of Nitrogen, Phosphorus, and Sediment from the Susquehanna River Basin, USA: Analysis of Decadal-scale Trends and Sub-basin Mass Balances

Qian Zhang ^a, William P. Ball ^{a,b}, Douglas L. Moyer ^c

^a Johns Hopkins University, Geography and Environmental Engineering

^b Chesapeake Research Consortium

^c U.S. Geological Survey, Virginia Water Science Center

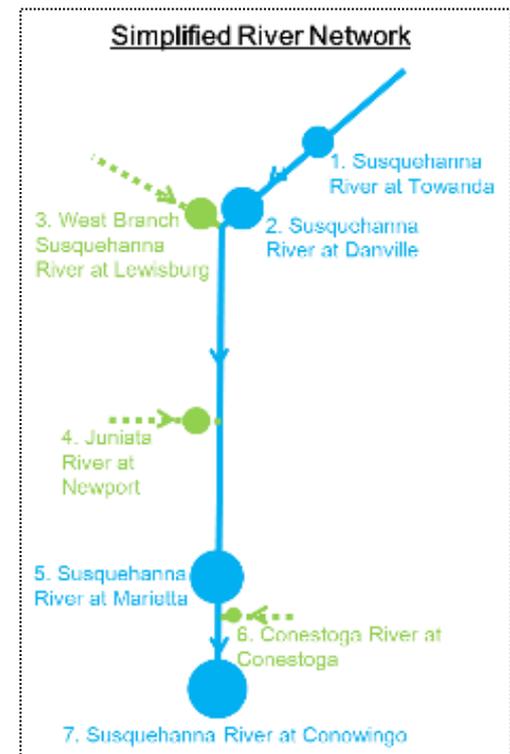
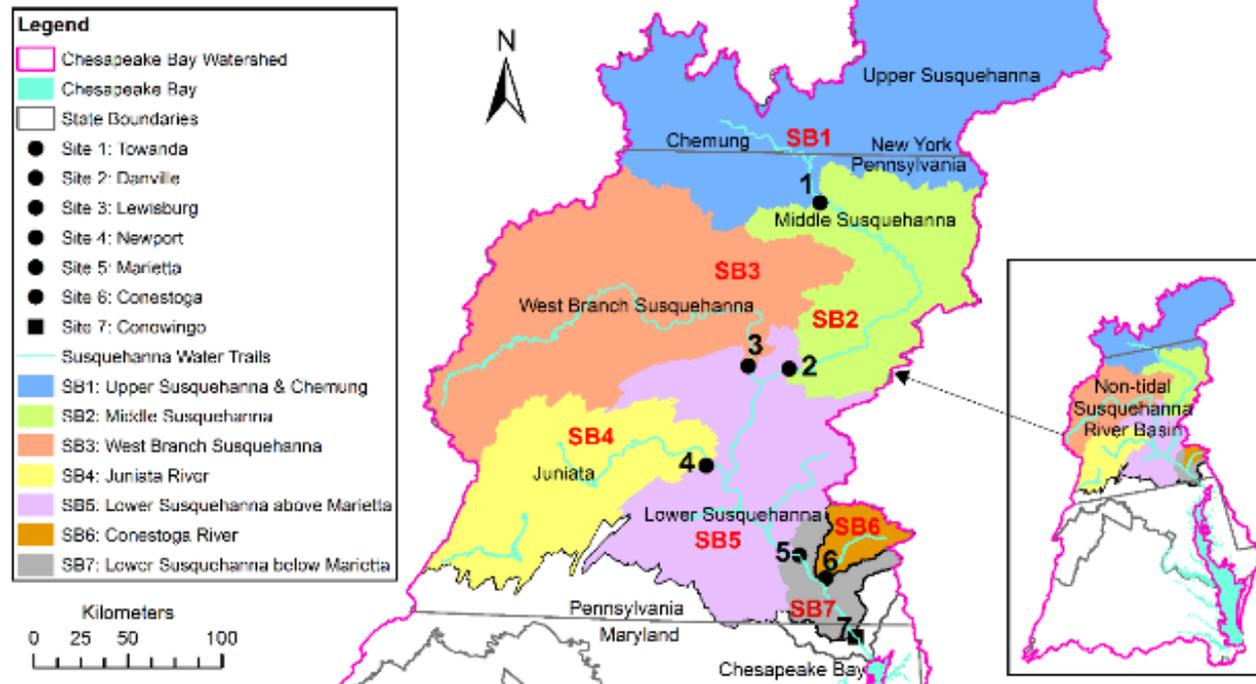
**GSA Annual Meeting 2015
Baltimore MD, Nov 1**

Objective

Objective: to better understand the historical progress in managing N, P, and SS exports from the Susquehanna River Basin (SRB) -- the largest tributary to Chesapeake Bay.

- **Trend analysis:** N, P, and SS loadings at seven sites.
- **Mass-balance analysis:** spatial budgets of major sub-basins, with focus on streamflow and land use effects on export.

Susquehanna River Basin



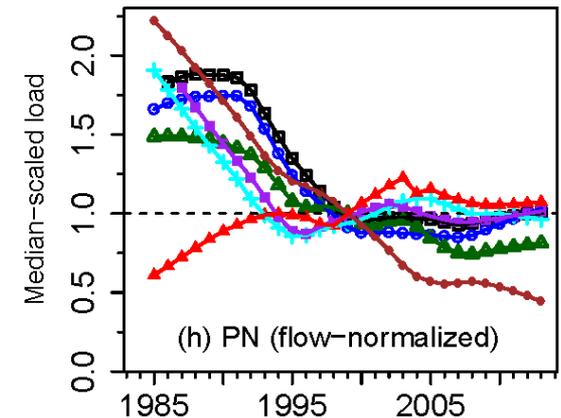
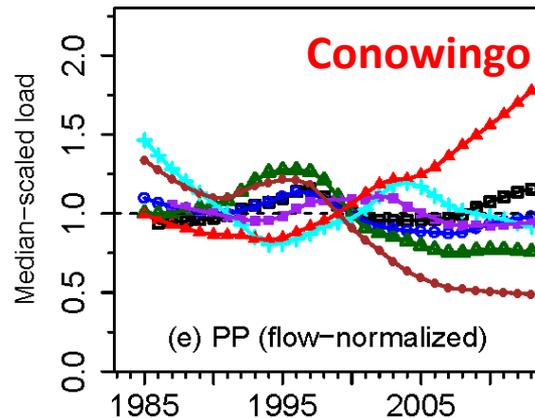
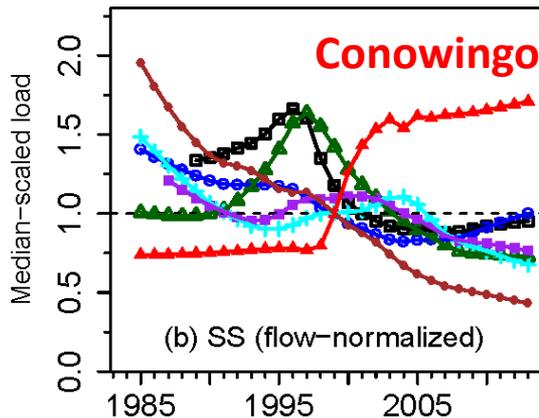
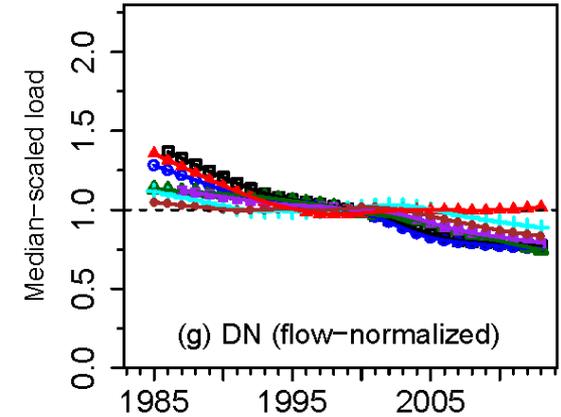
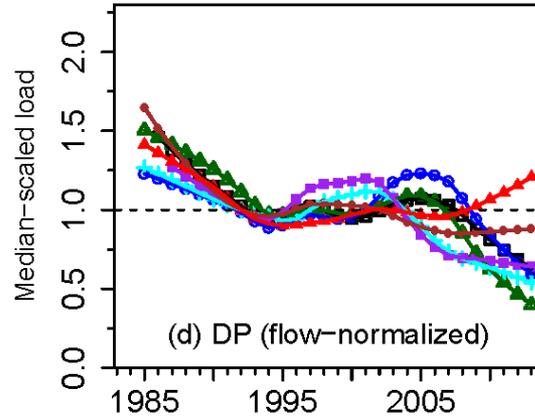
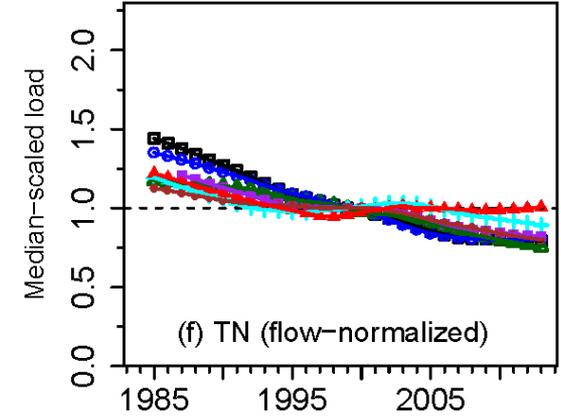
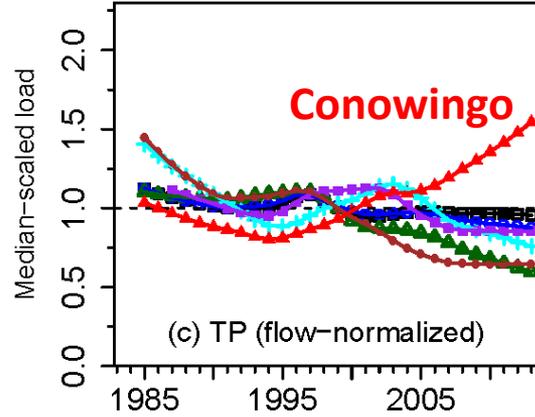
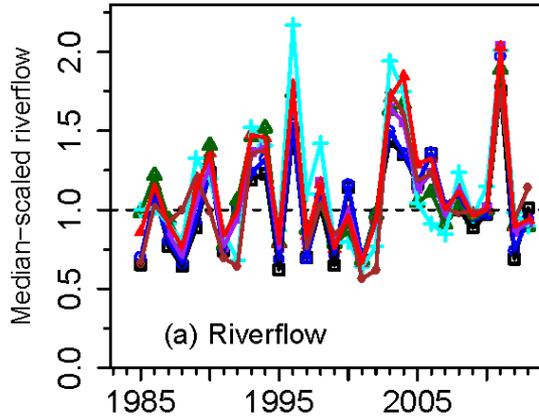
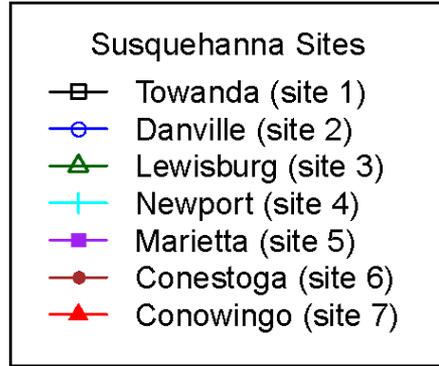
At each site

- **Monitoring data**: 25-40 sampled days per year on average; 1980s-2013.
- **WRTDS** (“Weighted Regressions on Time, Discharge, and Season”, Hirsch *et al.*, 2010) was run to estimate the daily **true-condition** and **flow-normalized** loadings* for each constituent, *i.e.*, **SS, TP, TN, DP, and DN**. [*“Loading” = riverine fluxes, unless otherwise stated.]
- Particulate nutrients: $PP = TP - DP$; $PN = TN - DN$.
- Annual loading / drainage area = annual riverine yield ($\text{kg km}^{-2} \text{ yr}^{-1}$).

For each sub-basin

- **Riverine input load**: flux entering the river reach = flux passing the monitoring site at the *upstream limit* of the river reach + flux entering from tributaries to that river reach (if any). [**≠ Watershed source input!**]
- **Riverine output load**: flux passing the site at the *downstream limit* of the river reach.
- **Net storage** = riverine input loading – riverine output loading.

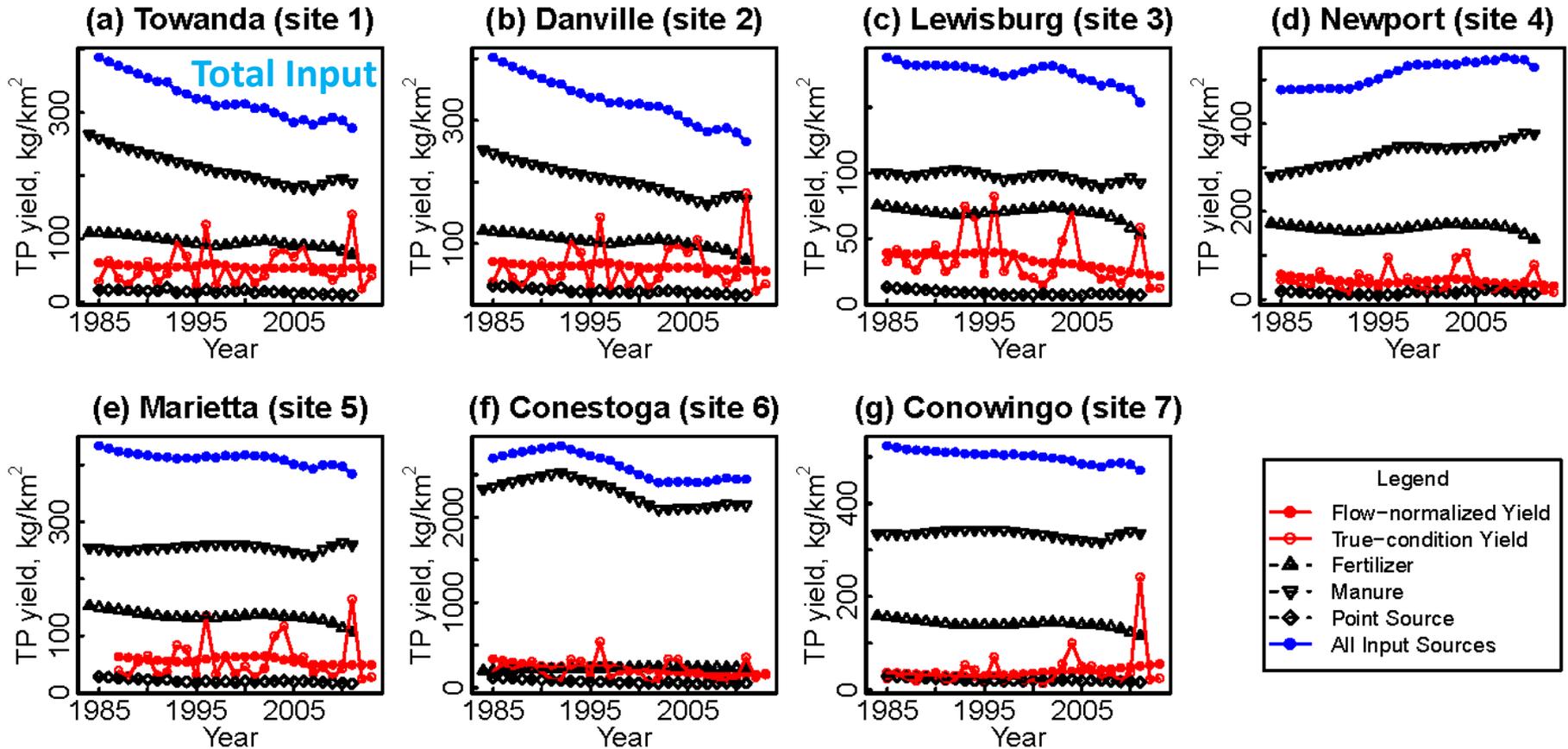
Q1: What have been the general patterns of long-term trends in riverine nutrient/sediment loadings? In particular, have trends been consistent (a) across the monitoring locations and (b) between dissolved and particulate species?



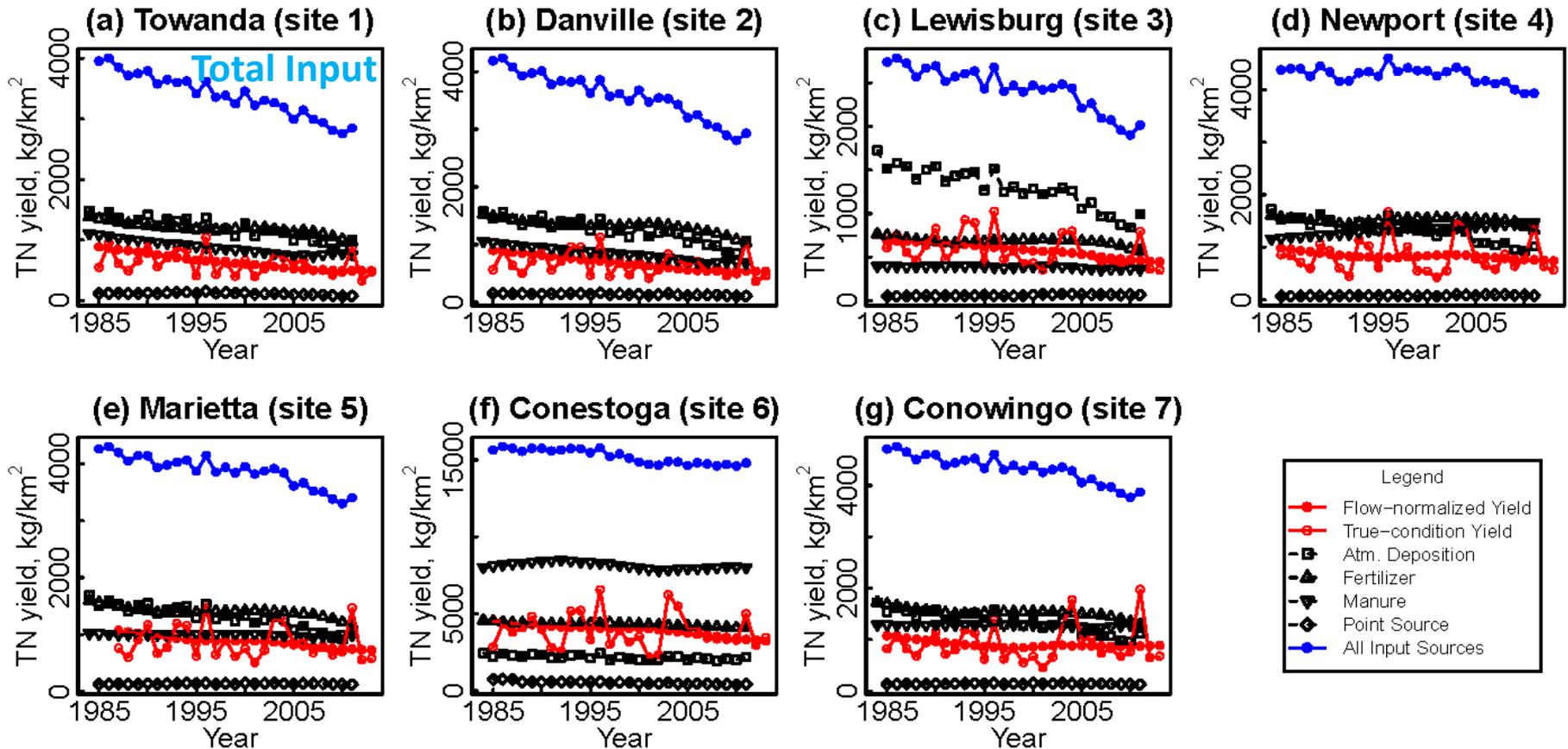
- **Upstream declines:** general declines at all sites upstream of the Conowingo Reservoir → indicating watershed-wide progress after decades of continued controls on various source inputs, including:
 - WWTP upgrade since the 1980s (CBP, 1998),
 - P-detergent ban in PA in 1990 (Litke, 1999),
 - P-based nutrient management (Weld *et al.*, 2002),
 - PA's Nutrient Management Act in 1993 and subsequent amendments for regulating CAOs/CAFOs (PA DEP, 2004),
 - Co-benefit of Clean Air Act on NO_x (Linker *et al.*, 2013).
- **Conowingo rises (SS, TP, PP, PN):** primarily driven by decreased reservoir trapping capacity (Hirsch, 2012; Zhang *et al.*, 2013) and possibly associated effects on biogeochemical transformations due to reduced residence time within the reservoir system.

Q2: What have been the general changes in watershed source inputs* and how have their magnitudes compared with those in riverine loadings?

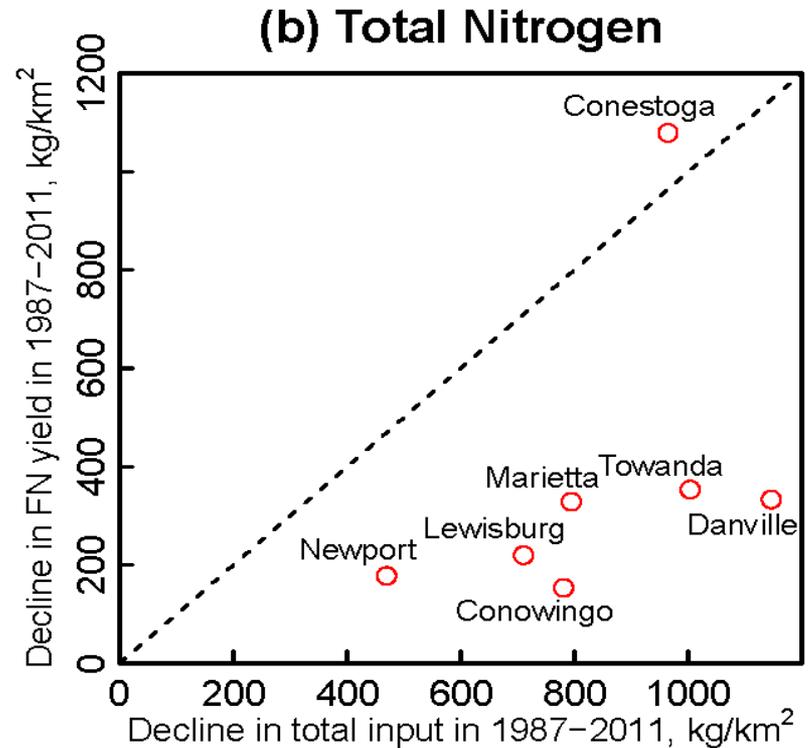
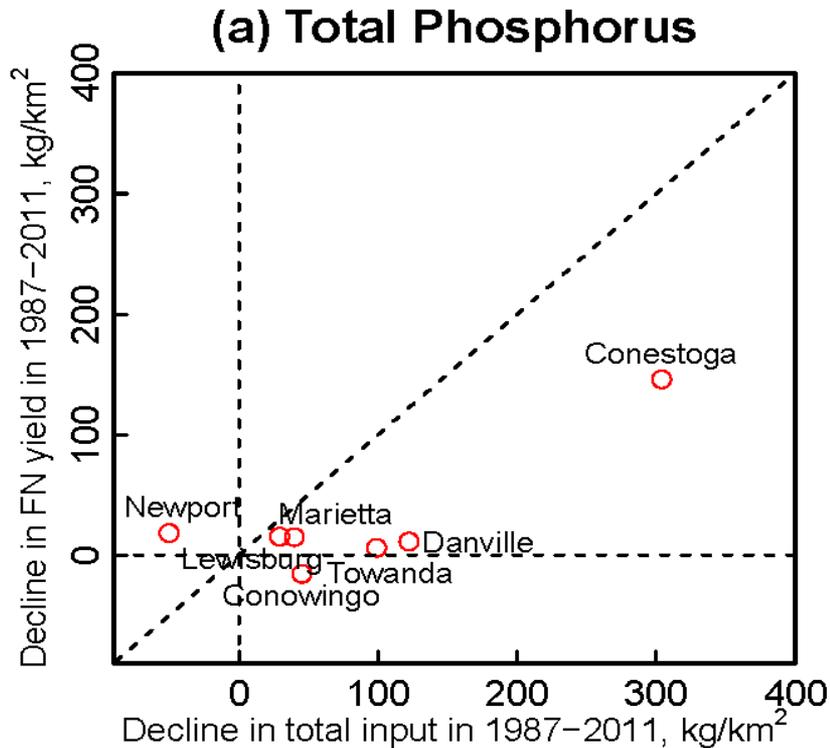
(*Based on input data from the Chesapeake Community Modeling Program, as provided by **Gary Shenk, Guido Yactayo, and Gopal Bhatt.**)



- **TP total source input:** declines at all sites except Newport, with the largest decline at **Conestoga**.
- **Individual sources:** negative Δ for 17 of 21 source-site combinations; dominated by manure and fertilizer reduction.
- **FN riverine yield:** negative Δ at all sites except Conowingo, with the largest decline at **Conestoga**.



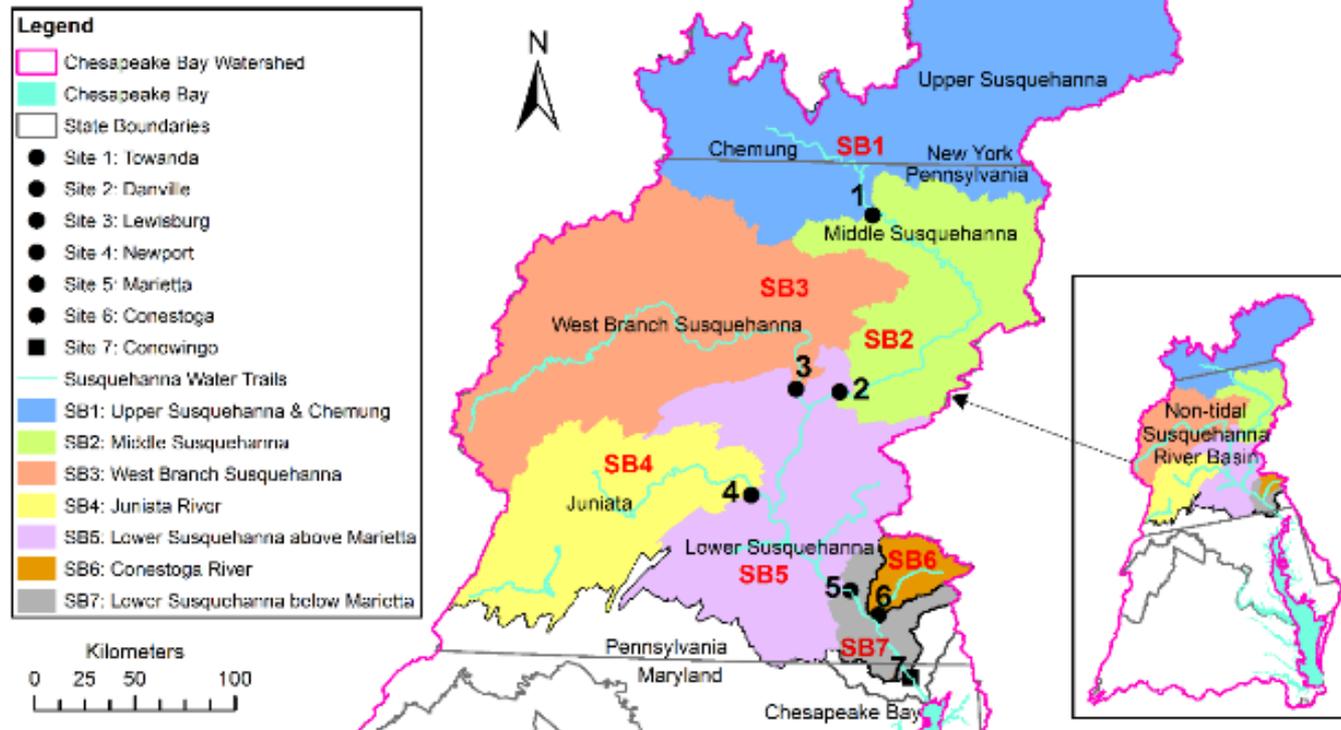
- **TN total source input:** declines at all sites, with the largest decline at Danville.
- **Individual sources:** negative Δ for 25 of 28 source-site combinations; dominated by atm. deposition (AD) reduction at all sites except Conestoga.
- **FN riverine yield:** negative Δ at all sites, with the largest decline at Conestoga.

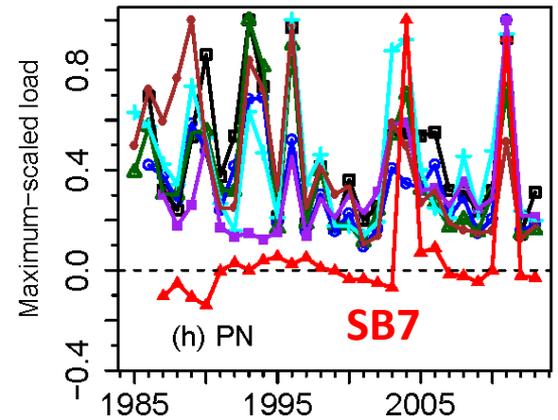
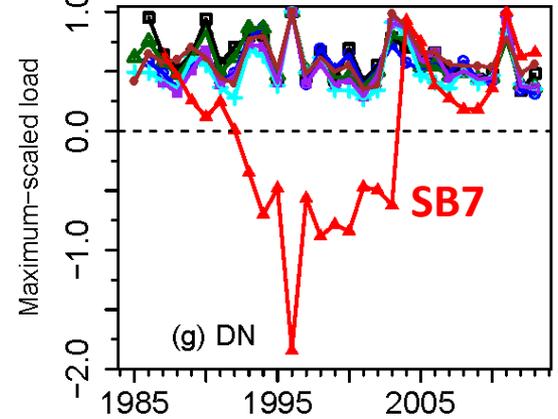
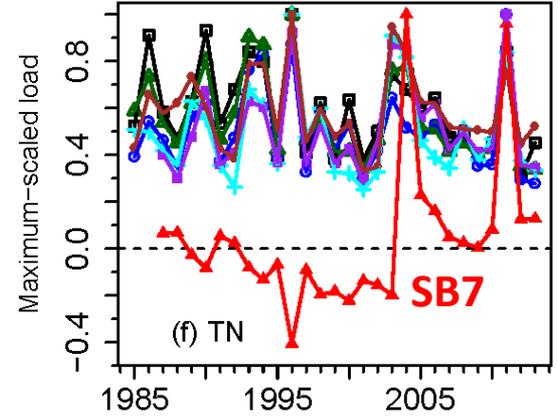
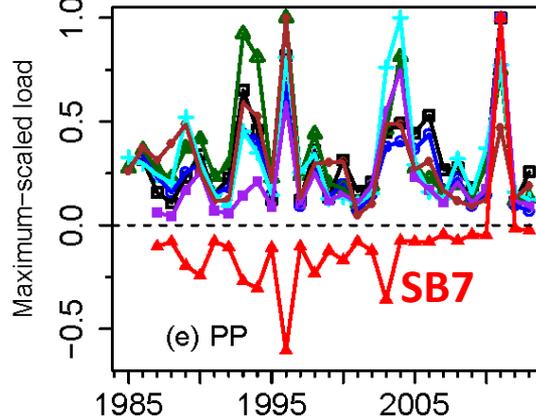
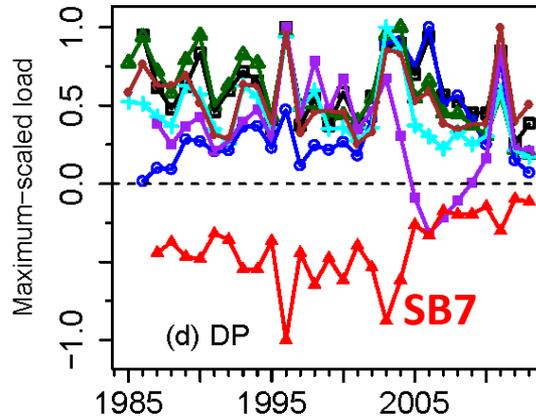
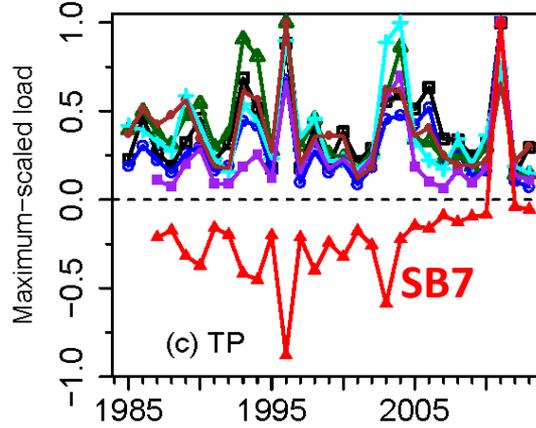
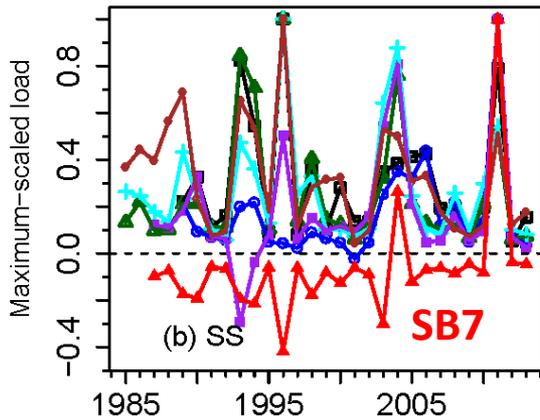
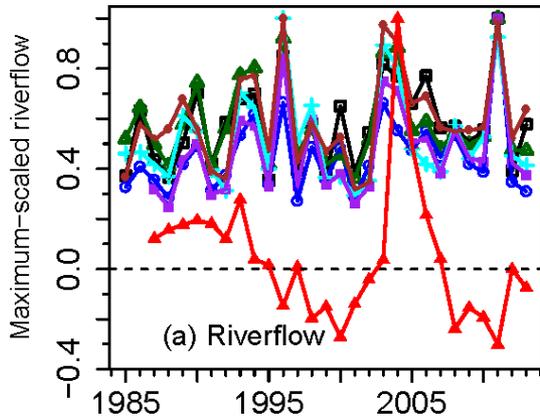
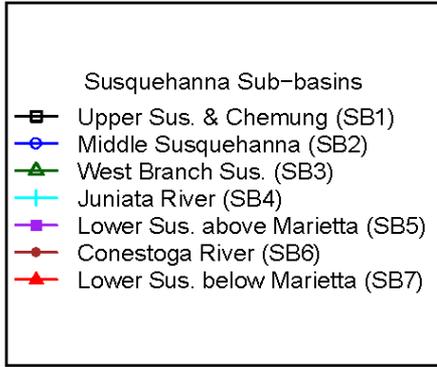


- $\Delta_{FN-Yield}/\Delta_{Input} > 0$ (12 of 14 cases) \rightarrow riverine yield has declined in response to reductions in watershed source input at different parts of the SRB.
- $\Delta_{FN-Yield}/\Delta_{Input} < 1.0$ (13 of 14 cases) \rightarrow likely continual contribution from legacy surface and sub-surface stores, as observed elsewhere and recognized as “**biogeochemical stationarity**” (Basu *et al.*, 2010; Thompson *et al.*, 2011).

Q3: Which sub-basins have been net sources (or storages) of loadings and what have been the role of streamflow on constituent export?

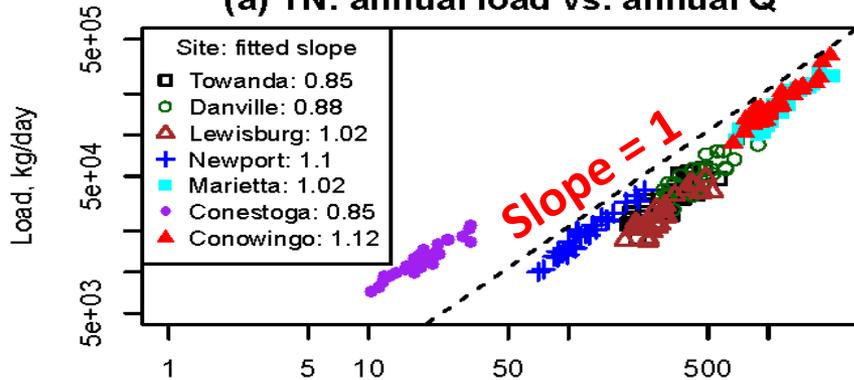
Susquehanna River Basin



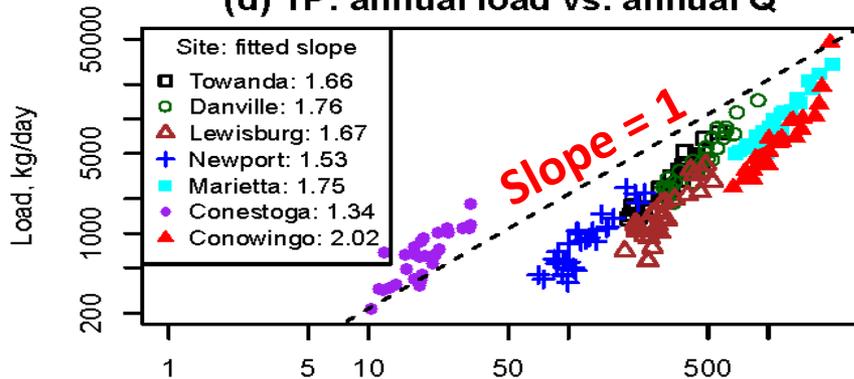


Sub-basins: Loading vs. Discharge Relationships

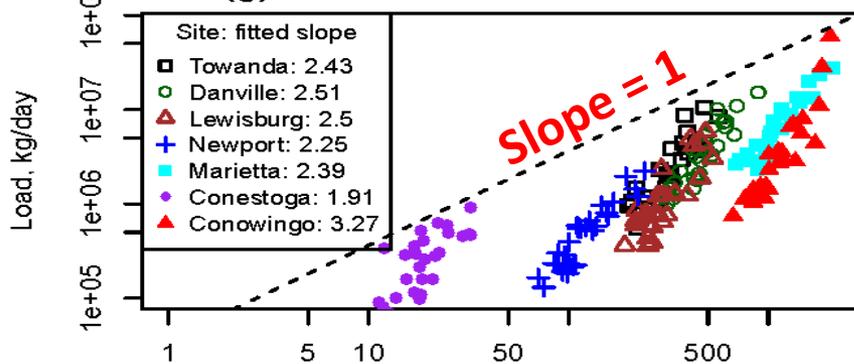
(a) TN: annual load vs. annual Q



(d) TP: annual load vs. annual Q

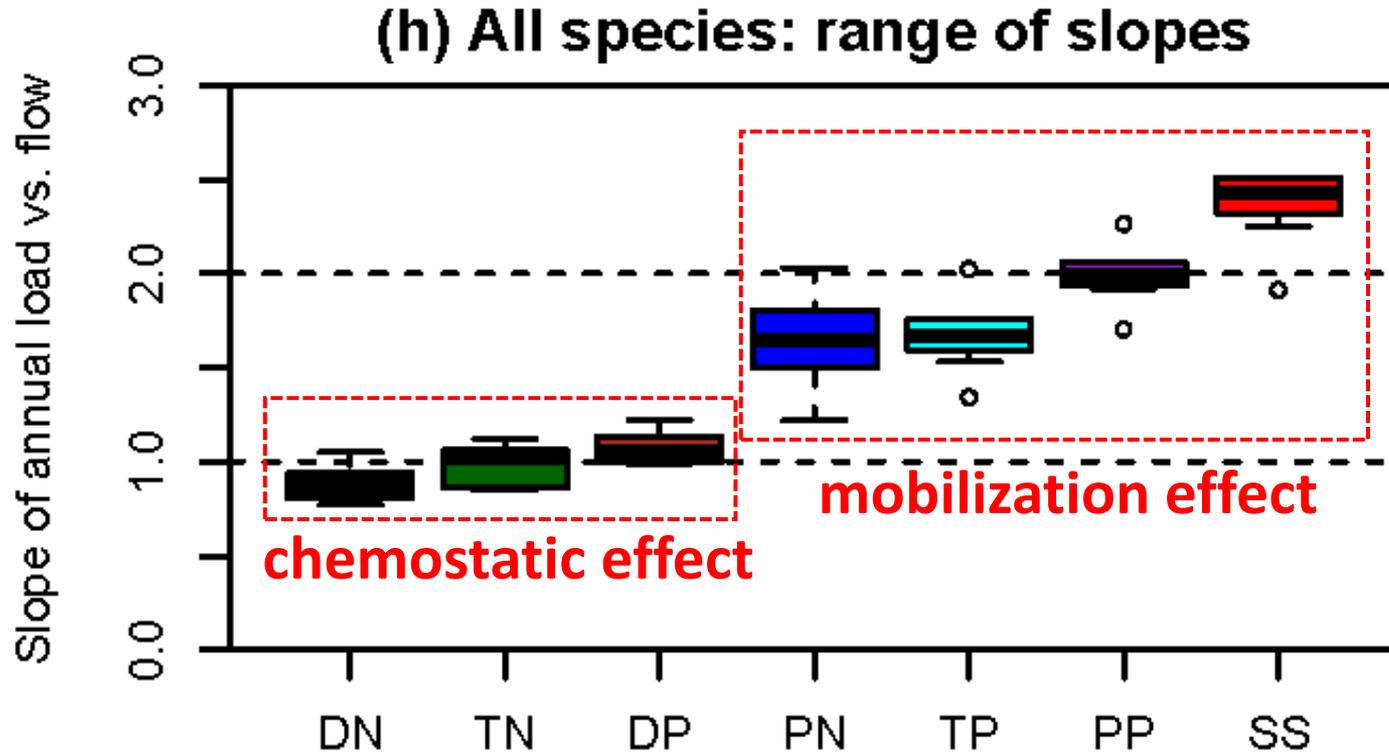


(g) SS: annual load vs. annual Q



Discharge, m³/s

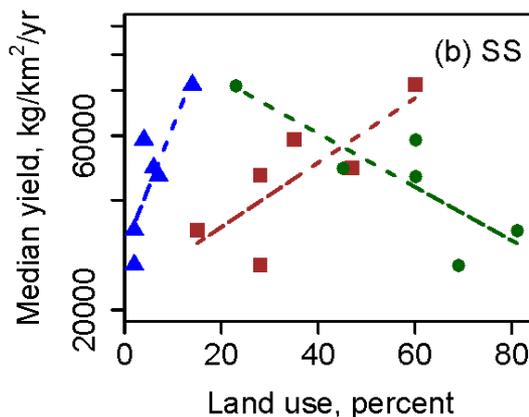
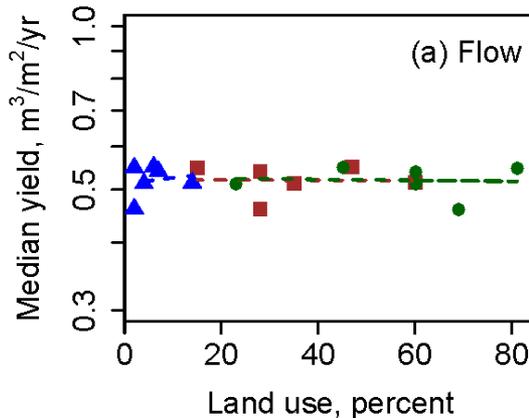
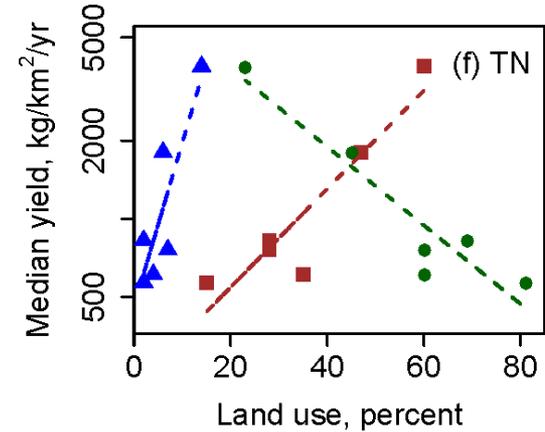
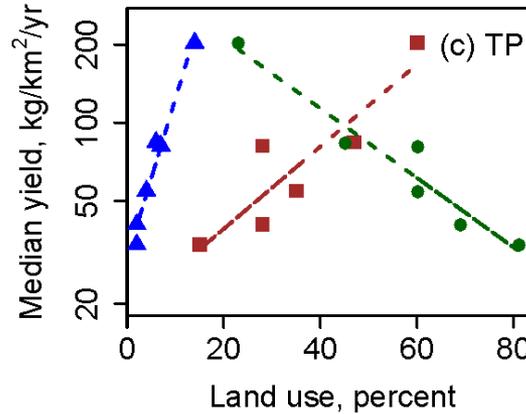
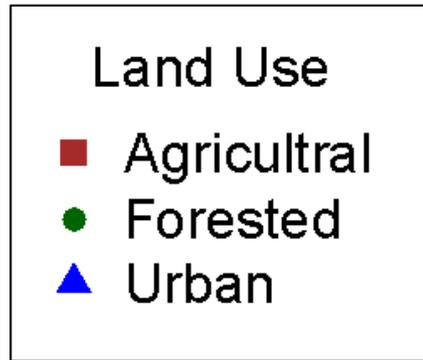
- **SB1-SB6:** concurrent peaks and troughs for Q and all constituents, showing a striking similarity with respect to the timing of significant export.
- **Hypothesis:** streamflow has been the principal factor controlling mass flux export in relative to other factors (*e.g.*, biogeochemical processes).
- **Validation:** strong log-linear relationships (p-value < 0.01) between annual loading and annual discharge for all species at all sites (DN, DP, PN, PP not shown) → Streamflow is a strong predictor of rates of N/P/SS export.
- **Literature:** consistent with studies on watersheds elsewhere (*e.g.*, Alvarez-Cobelas *et al.*, 2008; Basu *et al.*, 2010; Howarth *et al.*, 2006).



- **Implications:**

- ✓ exports not supply-limited;
- ✓ large storages despite decades of management efforts;
- ✓ “biogeochemical stationarity” (Basu *et al.*, 2010).

Q4: What have been the rankings of sub-basins with respect to nutrient/sediment yield and how have the rankings related to land use?



- **Streamflow:** annual flow yield is almost invariable with land use.
- **SS/TP/DP/PP/TN/DN/PN:** annual constituent yields correlate positively with the area fraction of **agricultural land and urban land** but negatively with that of **forested land** (due to lower source inputs and/or higher assimilation capacity).
- **Flow classes:** similar patterns observed with different flow classes (wet, dry, and average).
- **Literature:** consistent with findings discussed above and those on other watersheds (*e.g.*, Harris, 2001; Jordan *et al.*, 1997).

Conclusions

- N, P, and SS loads have declined at all Susquehanna sites upstream of Conowingo Reservoir.
- Smaller annual reductions in riverine yield than source input imply contribution of legacy sources.
- Riverflow has been a principal factor controlling rates of constituent export.
- Dissolved and particulate species show chemostasis and mobilization effects, respectively.
- Long-term yields of all species correlate positively with the fraction of agricultural or urban area but negatively with that of forested land.
- The study illustrates the value of maintaining long-term water-quality monitoring at multiple locations in watersheds.

References

- Alvarez-Cobelas, M., Angeler, D.G., Sánchez-Carrillo, S., 2008. *Environmental Pollution* 156, 261-269.
- Basu, N.B., Destouni, G., Jawitz, J.W., Thompson, S.E., Loukinova, N.V., Darracq, A., Zanardo, S., Yaeger, M., Sivapalan, M., Rinaldo, A., Rao, P.S.C., 2010. *Geophysical Research Letters* 37(23), L23404.
- Chesapeake Bay Program, 1998. Chesapeake Bay watershed model application and calculation of nutrient and sediment loadings, Appendix F: Point source loadings, p. 693.
- Gall, H.E., Park, J., Harman, C.J., Jawitz, J.W., Rao, P.S.C., 2013. *Landscape Ecology* 28(4), 651-664.
- Harris, G., 2001. *Marine and Freshwater Research* 52, 139-149.
- Hirsch, R.M., 2012. U.S. Geological Survey Scientific Investigations Report 2012–5185, p. 17.
- Hirsch, R.M., Moyer, D.L., Archfield, S.A., 2010. *Journal of the American Water Resources Association* 46(5), 857-880.
- Howarth, R.W., Swaney, D.P., Boyer, E.W., Marino, R., Jaworski, N., Goodale, C., 2006. *Biogeochemistry* 79(1), 163.
- Jordan, T.E., Correll, D.L., Weller, D.E., 1997. *Water Resources Research* 33(11), 2579-2579.
- Linker, L.C., Dennis, R., Shenk, G.W., Batiuk, R.A., Grimm, J., Wang, P., 2013. *Journal of the American Water Resources Association* 49(5), 1025-1041.
- Litke, D.W., 1999. US Geological Survey Water-Resources Investigations Report 99–4007, p. 43.
- Pennsylvania Department of Environmental Protection, 2004. Pennsylvania's Chesapeake Bay Tributary Strategy.
- Thompson, S.E., Basu, N.B., Lascurain Jr., J., Aubeneau, A., Rao, P.S.C., 2011. *Water Resources Research* 47, W00J05.
- Weld, J.L., Parsons, R.L., Beegle, D.B., Sharpley, A.N., Gburek, W.J., Clouser, W.R., 2002. *Journal of Soil and Water Conservation* 57(6), 448-454.
- Zhang, Q., Brady, D.C., Ball, W.P., 2013. *Science of the Total Environment* 452-453, 208-221.

Acknowledgements

- Maryland Sea Grant (NA10OAR4170072 and NA14OAR1470090)
- Maryland Water Resources Research Center (2015MD329B)
- National Science Foundation (CBET-1360415)
- Gary Shenk, Guido Yactayo, and Gopal Bhatt (Chesapeake Bay Program Office)
- Di Ha, Hengchen Wei (former JHU students)
- U.S. Geological Survey (NWIS data)
- Susquehanna River Basin Commission (SNAP data)



Questions?
Comments?